This document is confidential and is proprietary to the American Chemical Society and its authors. Do not copy or disclose without written permission. If you have received this item in error, notify the sender and delete all copies.

Solid and aqueous speciation of yttrium in passive remediation systems of acid mine drainage

Journal:	Environmental Science & Technology
Manuscript ID	es-2019-017956.R1
Manuscript Type:	Article
Date Submitted by the Author:	n/a
Complete List of Authors:	Lozano, Alba; Institute of Environmental Assessment and Water Research, Geosciences Fernandez-Martinez, Alejandro; CNRS, ISTerre Ayora, Carlos; Institute of Environmental Assessment and Water Research (IDÆA-CSIC), Department of Geosciences Di Tommaso, Devis; Queen Mary University of London, School of Biological and Chemical Sciences Poulain, Agnieszka; European Synchrotron Radiation Facility Rovezzi, Mauro; Univ. Grenoble Alpes, CNRS, IRD, Irstea, Météo France, OSUG, FAME Marini, Carlo; Consorcio para la Construccion Equipamiento y Explotacion del Laboratorio de Luz Sincrotron



1	Solid and aqueous speciation of yttrium in
2	passive remediation systems of acid mine
3	drainage
4	Alba Lozano ^{a,b*} , Alejandro Fernández-Martínez ^c , Carlos Ayora ^{a,b} , Devis Di Tommaso ^d ,
5	Agnieszka Poulain ^e , Mauro Rovezzi ^f , Carlo Marini ^g
6	^a Institute of Environmental Assessment and Water Research, (IDAEA-CSIC), Jordi
7	Girona 18-26, 08034 Barcelona, Spain
8	^b Grup de Mineralogia Aplicada i Geoquímica de Fluids, Departament de Cristal-
9	lografia, Mineralogia i Dipòsits Minerals, Facultat de Geologia, Universitat de
10	Barcelona (UB), C/Martí Franquès, S/N, Barcelona, Spain
11 12	^c Univ. Grenoble Alpes, Univ. Savoie Mont Blanc, CNRS, IRD, IFSTTAR, ISTerre, 38000 Grenoble, France
13	^d School of Biological Sciences, Queen Mary University of London, Mile End Road,
14	London, E1 4NS, UK
15	^e ESRF, The European Synchrotron, 71 Avenue des Martyrs, Grenoble 38000, France
16	^f Univ Grenoble Alpes, CNRS, IRD, Irstea, Météo France, OSUG, FAME, 38000
17	Grenoble, France
18	^g CELLS-ALBA Synchrotron Radiation Facility, Carrer de la Llum 2-26,08290
19	Cerdanyola del Valles, Barcelona, Spain
20	Corresponding author: Alba Lozano (<u>alba.lozano@idaea.csic.es</u>)

21 ABSTRACT

Yttrium belongs to the rare earth elements (REE) together with lanthanides and 22 23 scandium. REE are commonly used in modern technologies, and their limited supply has made it necessary to look for new alternative resources. Acid mine drainage (AMD) 24 is a potential resource since is moderately enriched in REE. In fact, in passive 25 remediation systems, which are implemented to minimize the environmental impacts of 26 AMD, REE are mainly retained in basaluminite, an aluminum hydroxysulfate 27 precipitate. In this study, the solid and liquid speciation and the local structure of 28 vttrium were studied in high sulfate aqueous solutions, basaluminite standards and 29 samples from remediation columns using synchrotron-based techniques and molecular 30 31 modeling. Pair distribution function (PDF) analyses and ab initio molecular dynamics density functional theory models of the yttrium sulfate solution show that the YSO_4^+ ion 32 pair forms a monodentate inner sphere complex. Extended X-ray absorption fine 33 34 structure (EXAFS) and PDF analyses show that Y is retained by basaluminite forming a monodentate inner-sphere surface complex at the aluminum hydroxide surface. EXAFS 35 of the column samples show that its signal is represented in more than 75% by the 36 signal of a basaluminite which YSO_4^+ is forming an inner-sphere complex. The atomic 37 vision of the REE configuration in AMD environments could help for a deeper research 38 39 of the REE recovery from waste generated in the AMD remediation systems.

40

41 INTRODUCTION

42 Yttrium belongs to Rare Earth Elements (REE) group, also composed by lanthanides and scandium. REE are necessary for the development of modern technologies, and 43 specifically, yttrium as important applications, for instance, in fluorescent lamps as 44 phosphors,¹ and in aircraft industry, used in the thermal barrier coatings for jet engines.² 45 The REE increasing demand but low worldwide supply has led to consider REE as 46 critical raw materials boosting searches for alternative resources, such as recycling used 47 stocks or identifying new geological sources of these elements. Because the REE 48 concentrations in acid mine drainage (AMD) are from one to two orders of magnitude 49 higher than the average concentrations in natural waters,³ it may be possible to perform 50 secondary REE recovery from precipitates from AMD neutralization in passive 51 remediation systems. The active systems were developed to minimize the environmental 52 impacts of AMD and they are used worldwide.^{4,5} However, due to the high water 53 content, sludge storage has substantial operational costs and environmental concerns.^{6,7} 54 In contrast, passive remediation systems, which have been developed extensively in 55 recent decades,⁸⁻¹¹ allow the AMD neutralization generating lower amounts of solid 56 waste precipitates. Ayora et al. documented nearly complete aqueous REE retention in 57 two laboratory columns simulating a disperse alkaline substrate (DAS), a passive 58 treatment already implemented in the field, for two highly acidic AMD (SW Spain).^{12,13} 59 The REE were scavenged by basaluminite, a mineral precipitated in the columns, which 60 also presented Y enrichment due to the higher vttrium concentration with respect to the 61 rest of REE in the two treated AMD. Basaluminite, an aluminum oxy-hydroxysulfate 62 $(Al_4(SO_4)(OH)_{10}, 5H_2O)$, precipitates in acidic environments as a consequence of the 63 64 natural attenuation of the AMD when mixing with more alkaline waters, or due to the induced neutralization of the acid waters, when the solution pH reaches ~ 4 .¹⁴ 65

Basaluminite is considered a nanomineral, with a short-range order, around 1 nm of
coherent domain size, which is described as layers of Al-octahedra with structural point
defects and with the presence of sulfate groups as outer-sphere complexes between the
Al layers.¹⁵

Similarly to the REE uptake by basaluminite in DAS treatments, Gammons et al., 70 reported the precipitation of hydrous aluminum oxides accompanied with a decrease in 71 REE concentration from AMD when mixed with natural water.¹⁶ Recently, the scavenge 72 of REE by basaluminite precipitates has been described as a sorption mechanism.¹⁷ 73 AMD is characterized to contain high loads of dissolved sulfate and the affinity of REE 74 to form aqueous species with sulfate is very high, the MSO₄⁺ aqueous complex being 75 the more abundant in AMD solutions.¹⁸ Sorption of dissolved REE from sulfate-rich 76 waters onto basaluminite is thus described as the sorption of the MSO₄⁺ aqueous 77 78 complex via ligand exchange with a surface site of basaluminite, forming a monodentate surface complex with the Al-octahedron as one proton is released.¹⁷ Here, 79 a structural description of the aqueous YSO_4^+ complex and of the local environment of 80 the surface complex formed upon adsorption onto basaluminite are reported. 81

Most of the previous studies reporting the adsorption mechanisms of REE using X-ray absorption spectroscopy and Extended X-ray Absorption Fine Structure (EXAFS) described the adsorption of free REE ions onto the surface of metal oxides.¹⁹⁻²² However, the AMD environments show a high concentration of dissolved sulfate, which presents high affinity for aqueous REE (M). Thus, the MSO₄⁺ aqueous complex is the major REE species in solution and explains that the sorption models of REE in AMD precipitates might be better explained in terms of MSO₄⁺ sorption.

The structure of the first hydration shell of lanthanides and yttrium has been well 89 characterized.²³⁻²⁸ The coordination number (CN) of the first hydration shell changes 90 from an average of CN ~ 9 oxygen atoms for light REE (LREE La to Pm) to CN ~ 8 91 oxygen atoms for heavy REE (HREE Tb to Lu), with intermediate values for medium 92 REE (MREE Pm to Gd).²⁷ This behavior is explained by the continuous decrease in 93 ionic radii with the atomic number.^{27,29,30} The ionic radius of yttrium is between that 94 those of Ho and Dy, so it is typically considered an HREE with a first hydration shell of 95 96 CN ~ 8 oxygen atoms. There are several studies describing aqueous complexes of REE with different ligands. REE easily forms inner-sphere complexes with carbonate and 97 phosphate in solution³¹⁻³³ whereas chloride and nitrate forms inner-sphere complexes at 98 concentrations higher than 10 M.^{23,25} In the case of sulfate, no information about the 99 geometry of MSO₄⁺ complexes has been reported so far despite their strong binding 100 affinity.18 101

The objective of this study is to elucidate the structure of Y adsorbed onto basaluminite. 102 Due to its chemical similarities with HREE, this allow us to assume similar structural 103 configuration for this sub-group. Moreover, this element was one of the most 104 105 concentrated in waste samples allowing performing X-ray absorption spectroscopy experiments. Since the YSO_4^+ aqueous complex is adsorbed onto the mineral,¹⁷ a 106 107 previous characterization of the geometry of the aqueous complex has been carried out. Finally a quantification of Y-species in basaluminite solids precipitated from AMD 108 treatments has been performed. Structural studies were performed using EXAFS and 109 pair distribution function (PDF) analyses of aqueous and solid samples combined with 110 *ab initio* molecular dynamics (AIMD) simulations of the aqueous YSO_4^+ complexes. 111

112 Two hypotheses are used to investigate the local structure of the aqueous YSO_4^+ ion 113 pair: (1) an outer-sphere complex, with water located between Y^{3+} and SO_4^{2-} , and (2) an inner sphere complex. In the latter case, two more hypotheses must be considered: (a) a monodentate complex, with one oxygen atom shared between the sulfate and the fist coordination sphere of Y^{3+} , and (b) a bidentate complex, with two oxygen atoms shared between the yttrium hydration sphere and the sulfate group.

118 Once the structure of the aqueous solution is fully described, different hypotheses have 119 been considered to interpret the YSO_4^+ surface complexation onto the Al-120 oxyhydroxysulfate: the formation of (1) a monodentate inner sphere complex, (2) a 121 bidentate mononuclear inner-sphere complex, (3) a bidentate binuclear inner sphere 122 complex.

Finally, to determine semi quantitatively the yttrium species in solids from column samples, linear combination fits were performed using the most representative references.

126 MATERIAL AND METHODS

127 Synthetic samples (standards)

128 Different basaluminites doped with Y were prepared for HEXS and EXAFS experiments. To ensure maximum Y incorporation into the solid, sorption and 129 coprecipitation experiments were performed at pH 6,¹⁷ avoiding higher pH values where 130 YCO_3^+ aqueous species could be present.³⁴ A basaluminite sample with sorbed YSO_4^+ 131 (B-YSO4-ads) was obtained from a suspension of 1g/L of solid in 0.11 Y(NO₃)₃·6H₂O 132 133 and 0.25 M Na₂SO₄ solution for 5 hours at a constant pH of 6. At these conditions, the major species in solution was YSO_4^+ , which is supposedly the aqueous Y specie sorbed 134 on the solid surface.¹⁷ A basaluminite sample coprecipitated with Y (B-Ycop) was 135 136 obtained using the basaluminite synthesis method adding 1 M Y(NO₃)₃·6H₂O in the

acid solution of 0.05 M Al₂(SO₄)₃·18H₂O (see Chapter S1 of the SI). Basaluminite and 137 nanoboehmite samples with sorbed Y^{3+} (B-Yads and A-Yads, respectively) were also 138 prepared from a suspension of 1 g/L of solid in 0.011-0.11 M Y(NO₃)₃·6H₂O solution 139 for 5 hours, and the pH was maintained at ~6 at room temperature, respectively (the 140 141 synthesis protocol is described in detail in Chapter S1 of the SI). At these conditions, the major species of Y is Y^{3+} . In all cases, the suspensions were centrifuged at 4500 rpm 142 for 15 minutes, the supernatant was filtered through 0.2 µm membranes and kept for 143 144 ICP analysis. The solids were rinsed several times with deionized water and oven dried for 48 hours at a maximum temperature of 40° C. 145

146 High-energy X-ray scattering (HEXS) and EXAFS experiments of aqueous solutions with Y and YSO₄⁺ were carried out. For the HEXS measurements, an aqueous solution 147 of free yttrium ions (Y-sol) was prepared with 0.1 M Y(NO₃)₃.6H₂O at pH 3.2, and a 148 149 solution containing YSO_4^+ as the major aqueous complex (YSO_4 -sol) were prepared by 150 mixing equal amounts of 0.2 M Y(NO₃)₃·6H₂O and 0.2 M Na₂SO₄ at pH 3. For the EXAFS measurements, a solution with the free Y^{3+} ion (Y-sol) with a composition of 151 0.01 M YCl₃· $6H_2O$ at pH 4.6, and the solution of the YSO₄⁺ complex (YSO₄-sol) with a 152 composition of 0.01 M YCl₃·6H₂O and 0.01 M Na₂SO₄ at pH 2.9, were prepared, 153 respectively. At these conditions, the major species in solution were Y^{3+} and YSO_4^+ 154 (Fig. S1). The vttrium concentrations of the references are listed in Table S1, and a 155 more comprehensive description of the syntheses is given in the Supporting Information 156 157 S1.

158 Waste samples from column treatments

159 Column samples were collected from two laboratory column experiments of Disperse Alkaline Substrate (DAS), as described in Ayora et al.¹² simulating the remediation 160 systems of the two AMD from the Iberian Pyrite Belt region (SW Spain). The columns 161 162 (9.6 cm inner diameter and 39 cm in height) consisted in a permeable mixture of grained limestone and wood shavings at a 1:1 weighed ratio. Samples W-MR-C1-4 and W-MR-163 C1-5 belonged to columns which treated AMD from Monte Romero mine and W-Alm-164 C3-8 and W-Alm-C3-9 samples belonged to columns which treated leaching of 165 Almagrera mine tailing dam. Y concentration in different AMDs at the inlet ranged 166 167 from 9 to 42 ppm, and sulfate concentrations 3.5 and 11.7 mg/L. Physical-chemical 168 parameters and water samples were collected every two weeks. Once the treatment finished after 18 weeks, the solid was divided in 2 cm thick slices and dried. The 169 170 partitioning of Al, Fe and REE was studied by performing a sequential extraction adapted from Torres and Auleda to the solid residues.³⁵ The results showed in Ayora et 171 al. concluded that REE were retained in basaluminite.¹⁰ The W-MR-C1-4, W-MR-C1-5, 172 W-Alm-C3-8 and W-Alm-C3-9 samples were selected from the depth were maximum 173 Al concentration was extracted from the solids, assuming the most enriched in 174 175 basaluminite solid slice. These waste samples were used for EXAFS measurements and 176 the Y concentrations for these samples are shown in Table S2.

177 Analytical techniques

Measurements of the pH values of the initial and final solutions were made with a Crison[®] glass electrode calibrated with buffer solutions of pH 2, 4 and 7. The Al and S concentrations were measured by inductively coupled plasma optical emission spectroscopy (ICP-OES), and the Y concentration was measured by inductively coupled plasma mass spectrometry (ICP-MS). The aqueous speciation and saturation index of the solid phases were calculated with the PHREEQC code³⁶ using the Donnee Thermoddem_V1.10 database³⁷ provided by the Bureau de Recherches Géologiques et Minières (BRGM). Details of the analytical procedures and the database for the REE speciation are described in the Supporting Information S2.

The HEXS measurements were performed at the ID31 beamline at the European 187 Synchrotron Radiation Facility (ESRF). The samples were loaded into 1.5 or 2 mm 188 Kapton capillaries, and the HEXS patterns were obtained using a Pilatus3 X CdTe 2M 189 detector. The data correction and generation of structure factors and pair distribution 190 functions were performed with the PDFgetX3 software.³⁸ PDF for aqueous solutions 191 were obtained by subtracting the scattering signal of pure water from the total scattering 192 signal of YSO4-sol sample and Y-sol sample, respectively. The differential pair 193 194 distribution functions (d-PDF) for B-Ycop sample were obtained by subtracting the PDF signal of a pure synthetic basaluminite from the PDF of B-Ycop sample. 195

196 The EXAFS measurements of the samples B-Yads, B-Ycop, A-Yads, Y-sol and YSO4sol were performed at the Y K-edge (17.038 keV) at the FAME beamline,³⁹ BM30B at 197 the ESRF. The EXAFS measurements of the B-YSO4 sample was performed at the Y 198 K-edge (17.038 keV) at CLÆSS beamline (ALBA synchrotron). Atomistic models of 199 YSO_4^+ and Y^{3+} complexes adsorbed on a basaluminite surface with different initial 200 201 configurations were employed in the data analysis to extract the Y local order parameters, including the neighbor distances, R, coordination number, N, and Debve-202 Waller factors, σ^2 . The proportion of Y species present in each waste sample from 203 204 column treatments was obtained by linear combination fitting (LCF) of the EXAFS spectra using the most representative reference compounds as the basis set. The 205

experimental setup and the details of the data analysis are given in the SupportingInformation S3.

The Ab Initio Molecular Dynamics (AIMD) simulations of an aqueous solution of 208 vttrium(III) sulfate were carried out on a Y^{3+} -SO₄²⁻-OH⁻ system embedded in a box of 209 61 water molecules. The simulations were conducted with the electronic structure code 210 CP2K/Quickstep code, version 5.1.40 CP2K implements density functional theory 211 212 (DFT) based on a hybrid Gaussian plane wave. The revPBE generalized gradient approximations for the exchange and correlation terms were used together with the 213 general dispersion correction termed DFT-D3.^{41,42} The simulations were carried out for 214 215 15 ps in the canonical (constant NVT) ensemble to maintain the average temperature at T = 300 K. An in-house Python code using DiffPy-CMI was developed to convert the 216 AIMD trajectories to PDF signals.⁴³ Details of the MD simulations are described in S4. 217

218 RESULTS AND DISCUSSION

219 Geometry of the aqueous complex YSO₄⁺

220 The experimental (YSO4-sol, Y-sol) and theoretical (YSO4-calc) PDFs of the Y-221 containing aqueous solutions, 0.1 M YNO₃ and YSO₄, are shown in Fig. 1. Background subtraction of the aqueous samples was performed using a deionized water background, 222 223 which implies that the experimental PDFs shown here are differential PDFs (water subtracted). The background scale factor was adjusted to minimize the O-O correlation 224 (~2.8 Å) in the PDFs. All spectra show an intense peak at 2.37 Å, which corresponds to 225 the Y-O interatomic distances to oxygen in the first coordination shell. The 226 experimental PDF of the YSO4-sol sample shows a small and sharp peak at 1.45 Å that 227 corresponds unequivocally to the S-O interatomic distance in sulfate, and another 228

smaller peak at ~3.6 Å, which is absent in Y-sol sample. The DFT-based PDFs obtained from AIMD simulations of the YSO_4^+ aqueous complex also reproduced this latter peak. Analyses of the AIMD trajectories show that the 3.6 Å distance between Y and S is consistent with the formation of a monodentate inner sphere complex (one shared oxygen) between the coordination shells of the sulfur and yttrium.



Figure 1. (A) Top: Experimental PDFs of YSO4-sol and Y-sol samples. Bottom: Simulated (AIMD) PDF (YSO4-calc) and partial PDFs of an YSO₄⁺ aqueous complex. (B) Fourier-filtered signal from 1.8 to 4.2 Å for the EXAFS data. (C) EXAFS FT amplitude functions of the YSO4-sol sample. Black lines: experimental; red lines: fits. Simulated (AIMD) PDF and partial PDFs have been multiplied for visualization purposes: YSO₄-cal (×3), Y-S (×5) and Y-O and S-O (×2). Dashed lines indicate the position of the Y-O, Y-S and S-O bonds in YSO4-sol sample.

Fits of the EXAFS spectra of the aqueous solutions were performed using atomistic models from the AIMD simulations as initial structural models for the generation of the

scattering paths. The results yielded an average Y-O distance of 2.37 ± 0.02 Å for the 244 Y-sol sample and 2.38 \pm 0.01 Å for YSO4-sol (Table 1). The intense peak in the FT 245 function at 2.38 Å for the two aqueous references represents the first solvation shell, 246 and its asymmetry reveals a distribution of Y-O interatomic distances, as reported 247 previously by Lindqvist et al.²⁴ The coordination numbers (CN) were 7.6 \pm 1.9 and 7.9 \pm 248 0.9 for Y-sol and YSO4-sol, respectively, which are close to the expected value of 8 for 249 HREE. A second shell is visible in the FT of the YSO4-sol sample. It was fitted with a 250 251 Y-S distance of 3.50 ± 0.04 Å, similar to that observed in the PDF of the same sample, and a CN = 1. This corresponds to a monodentate coordination, matching the geometry 252 obtained from the AIMD simulations. The coordination number CN = 1 was kept fixed 253 in the EXAFS fitting to increase the stability of the fitting procedure, minimizing the 254 number of parameters and excessive correlations between them. A multiple scattering 255 256 path (Y-S-O) was included and was found to be relevant and to improve the fit. The fitting parameters matched both the PDF and modeling results and indicated that the 257 YSO_4^+ aqueous species forms a monodentate inner-sphere complex. 258

Table 1. Modeling parameters of the Y K-edge EXAFS spectra of the 0.1 M YSO₄
aqueous solution (the error is expressed in the parentheses after the last digit).

Sample	Neighbor	path	Ν	σ^2	ΔE_0	R (Å)	Var.	χ_{ν}^{2}
Y sol	1 st shell	Y-O	7(2)	0.005(3)	-2(3)	2.37(2)	4	354
Y-SO4 sol	1 st shell	Y-O ₁	7.9(9)	0.007(2)		2.38(1)	8	498
	2 nd shell	Y-S	$1_{\rm fixed}$	0.003(6)	-1(1)	3.50(4)		
	3 rd shell	Y-S-O	$1_{\rm fixed}$	0.015(19)		3.18(10)		

The determined Y-O first shell coordination numbers and interatomic distances are 261 consistent with previous reports describing yttrium coordination shells with N = 8 and 262 Y-O distances of 2.36 Å.^{23,24,28} Likewise aqueous REE carbonate and phosphate 263 complexes,³¹⁻³³ aqueous Y-SO4 ion-pairs form inner-sphere complexes. This result 264 contrasts with other ligands such as chloride, which hardly forms inner sphere 265 complexes at similar concentrations to those used in this study.²³ Only the formation of 266 monodentate complexation between Dy and NO₃ with a distance of 3.57 Å has been 267 reported using MD simulations.²⁶ 268

269 Local order of yttrium adsorbed into basaluminite

HEXS measurements were performed on a pure basaluminite sample (B-pure) and on a 270 sample of basaluminite coprecipitated with Y (B-Ycop). The PDFs are shown in Fig. 2 271 272 with the differential PDF obtained by subtracting the pure sample from the 273 coprecipitated one. This approach has been previously used to examine the local order of different anions incorporated in poorly crystalline minerals, such as schwertmannite 274 275 and basaluminite, and the desorption of sulfate from basaluminite when ageing at circumneutral pH values.^{15,46} The PDFs of the two samples are similar and show 276 characteristic basaluminite peaks that correspond to the known interatomic distances (S-277 O bond at ~1.46 Å, Al-O bonds at 1.88, 4, 4.8 and 6 Å and Al-Al bonds at 3 Å).^{15,45} The 278 differential PDF shows a main peak at 2.37 Å, which corresponds to the Y-O distance. 279 280 This peak matches the distance from Y to the first solvation shell obtained from the EXAFS and PDF of the Y^{3+} and YSO_4^+ aqueous solutions (Fig. 1). Another peak is 281 present at 1.7 Å, which is tentatively assigned to a new Al-O distance of the Al-282 283 octahedra, due to probable distortions of the basaluminite. The existence of deformations in the structures of mineral nanoparticles upon oxyanion adsorption has 284 been previously observed.^{44,46} A smaller peak appears at 3.48 Å, which can be 285

tentatively assigned to a Y-S distance, though it is shorter than the Y-S distance reported for the YSO_4^+ aqueous complex (Table 2 and Fig. 1A). This peak may also correspond to a Y-Al interatomic distance, which would be consistent with the formation of an inner-sphere complex, as suggested by Lozano et al.¹⁷ This hypothesis will be tested using different atomistic models for the fitting of the EXAFS data. Other peaks at longer distances are difficult to assign in the absence of a more detailed structural model.



293

Figure 2. PDFs of basaluminite coprecipitated in the presence of Y (B-Ycop), pure basaluminite (B-pure) and differential PDF (d-PDF). The d-PDF spectrum has been amplified (×3) for visualization purposes.

The k³-weighted EXAFS spectra for the solid references B-YSO4-ads, B-Y-ads, B-Ycop and A-Y-ads (Fig. 3) are indistinguishable and differ from the aqueous Y-sol and YSO4-sol spectra in that a slight oscillation becomes visible at 5-6 Å⁻¹. In the B-YSO4ads the dominant Y species in solution is the YSO_4^+ complex. However, in B-Y-ads, B-Ycop and A-Yads, the dominant Y species is Y^{3+} . Still, these four samples show similar spectra. Since the common feature in these three samples is the presence of an Al adsorbent, and this feature agrees with that in the B-YSO4-ads EXAFS spectrum, we

hypothesize that the higher frequency observed at 5-6 \AA^{-1} (arrow in Fig. 3) may be 304 related to the presence of an Al backscatterer, rather than to a signal from an close 305 sulfate group. Both Al and S are light atoms with similar atomic numbers (Z = 13 and 306 16, respectively), which makes their backscattering functions difficult to distinguish via 307 fitting of the EXAFS signal. These observations lead us to assume that the EXAFS 308 measurements have poor sensitivity to the presence of Y-S bonds in this system. 309 310 However, a contribution from a sulfate shell for the two first samples cannot be ruled 311 out.



312 [

Figure 3. k³-weighted EXAFS (A) and FT amplitude functions (B) for four waste samples from column treatments W-MR-C1-3, W-MR-C1-4, W-Alm-C3-8, and W-Alm-C3-9 (upper part) and solid standards (basaluminite sorbed with YSO4. (B-YSO4), basaluminite sorbed and coprecipitated with Y: B-Yads and B-Ycop, respectively) and

aqueous solution (free ion and sulfate complex: Y-sol and YSO4-sol, respectively)
(bottom part). The dashed lines in the EXAFS signals of the column samples represent
LCF with B-Yads (basaluminite with sorbed yttrium) and YSO4-sol (solution of Y with
SO₄) standards as the most representative references (results in Table 3). The arrows
indicate a frequency present in the solid standards.

Based on the previous information of the aqueous species and on the thermodynamic model of REE sorption,¹⁷ the B-YSO4-ads EXAFS signal was fitted with three models of different geometries of the YSO_4^+ aqueous complex adsorbed onto basaluminite surfaces via inner-sphere complexes (Fig. 4), to obtain an atomistic model representation: (A) a monodentate surface complex; (B) a bidentate mononuclear surface complex and (C) a bidentate binuclear surface complex.



Figure 4. Atomistic representations of the three models of YSO4 aqueous complex adsorbed on the basaluminite-water interface. The different atomic positions of YSO4 to octahedral-Al are used to fit the EXAFS signal of the B-YSO4 sample. The three

models show different inner-sphere surface complexes: (A) monodentate, (B) bidentatemononuclear and (C) bidentate binuclear.

The EXAFS fits for B-YSO4-ads for the different models and the calculated parameters 334 are given in Fig. 5 and Table 2, respectively. The first model represents a monodentate 335 336 inner-sphere surface complex. The coordination number of the shells has been fixed based on information from another sources: The major aqueous species in solution was 337 YSO_4^+ , which, as supported by PDF and DFT data, forms a monodentate aqueous 338 complex. Besides, the thermodynamic model of REE sorption on basaluminite indicated 339 monodentate complexation of the YSO_4^+ aqueous complex.¹⁷ With this information the 340 fit was performed assuming S and Al coordination numbers equal to 1 and 2, 341 342 respectively, as it is shown in Fig. 4A. Besides, two oxygen paths were considered due to the Y-O length variation improving the fit. Thus, the fit yielded a first coordination 343 344 shell with a Y-O distance range between 2.32 and 2.47 ± 0.02 Å with a fix coordination number of 8. The second shell was fitted yielding a Y-Al distance at 3.52 Å with a fixed 345 coordination number of N = 2, and a Y-S distance at 3.33 with a fixed number of N = 1. 346 This result is consistent with the monodentate inner-sphere complex of Y with one 347 oxygen shared with two surface Al sites (Fig. 4A). The second model, the bidentate 348 349 binuclear inner-sphere complex configuration, gave poorer results without structural significance and with higher χ_v^2 values (Table 2), so it was discarded. The third model 350 was the bidentate binuclear, the first shell was fitted with a coordination number of $10 \pm$ 351 352 1.31 and a distance of 2.33 Å. The second shell was fitted with a S path yielding an Y-S distance of 3.34 Å with a fix coordination number of N = 1, similar to that for the 353 monodentate hypothesis, and with an Al path with a fixed coordination number of N = 2354 giving a distance a bit longer, 3.92 Å. 355

Overall, the monodentate inner-sphere surface complex gave the best fit results, which 356 was in accordance to the thermodynamic sorption model. Moreover, the bidentate 357 binuclear inner-sphere surface complexes yielded also a good fit with similar χ_v^2 values 358 than the monodentate hypothesis, making this configuration also possible (Fig. 4C). In 359 both fits, the distances between Y-S are similar, 3.33, 3.34 Å, which is also observed in 360 the differential PDF of the B-Ycop reference. From the EXAFS data, the presence of a 361 bidentate binuclear surface complex cannot be ruled out. However, the thermodynamic 362 363 modeling and the EXAFS data both agree on the monodentate surface complex being the one present at the basaluminite-water interface. 364



366	Figure 5. (A) The k^3 -weighted EXAFS spectra at the Y K-edge of the basaluminite with
367	YSO4 sorbed (B-YSO4 reference) and (B) its Fourier transform amplitude. The
368	experimental and fitted curves are shown in black and red, respectively.
369	Table 2. Results of the EXAFS fits for B-YSO4 reference. Model 1: monodentate inner

370 sphere. Model 2: bidentate mononuclear inner sphere. Model 3: bidentate binuclear

inner sphere. Var. indicates independent variables. The best model is indicated with a

star (*). The error is expressed in the parentheses after the last digit.

Model	Neighbor	path	Ν	σ^2	ΔE_0	R	Var.	χ_{v}^{2}
1*	1 st shell	$Y-O_1$	$6_{\rm fix}$	0.007(1)	-3(2)	2.32(1)	9	338
		Y-O ₂	2_{fix}	0.004(8)	-3(2)	2.47(2)		
	2 nd shell	Y-S	$1_{\rm fix}$	0.008(5)	-3(2)	3.33(4)		
		Y-Al	2_{fix}	0.012(5)	-3(2)	3.52(5)		
2	1 st shell	Y-O	10(1)	0.012(2)	-6(2)	2.33 (1)	8	402
	2 nd shell	Y-S	$1_{\rm fix}$	0.009(5)	-6(2)	3.88 (5)		
		Y-Al	1_{fix}	0.006(3)	-6(2)	3.42 (3)		
3	1 st shell	Y-O	10(1)	0.013(1)	-4(1)	2.33(1)	8	334
	2 nd shell	Y-S	$1_{\rm fix}$	0.008(3)	-4(1)	3.34(3)		
		Y-Al	2_{fix}	0.011(5)	-4(1)	3.92(4)		

373

374 The EXAFS of the B-YSO4-ads standard was also fitted with the same three atomistic models presented in Fig. 4 but without sulfate included (Fig. S2) to analyze the 375 376 sensitivity of the fit to the presence of sulfate. The Al-O distances obtained with these 377 new fits (Table S3) were almost identical to those calculated when including sulfate shells (Table 2), with the exception in the monodentate inner sphere configuration, 378 which was a bit longer. Thus, the presence of Al is consistent in the two models, with 379 380 and without sulfate. However, the fits obtained with the sulfate group included showed lower values of the agreement factor (reduced chi-square) (compare Table 2 and Table 381 S3). This discussion highlights the difficulties of the EXAFS technique to discern the 382

presence of low-Z backscatterers, in particular in disordered systems such as the onepresent here.

These results are consistent with those presented by Rabung et al. where Gd and Lu 385 were shown to adsorb onto γ -Al₂O₃ to form inner sphere complexes.⁴⁷ Lu, which is 386 387 considered a HREE such as Y, was shown to form a monodentate surface complex with a first shell of 7 atoms and a Lu-O distance of 2.28 Å. The longer Y-O distance found 388 here for Y is consistent with its larger Y ionic radius. However, these authors could not 389 detect any Lu-Al bonds in their EXAFS data. Other sorption studies of Eu onto γ -Al₂O₃ 390 showed the formation of both monodentate and bidentate binuclear complexes.²⁰ In 391 392 contrast to this study and to the results of Rabung et al. on Al oxides, EXAFS studies of 393 lanthanides sorbed onto ferric oxides have shown a predominance of bidentate innersphere complexes.⁴⁷ Lu adsorbed onto ferrihydrite at pH \geq 5.5 yielded an Lu-O 394 coordination number of 7 at a distance of 2.30 Å, and a second coordination shell was 395 identified with a Lu-Fe interatomic distance at 3.38 Å. These EXAFS results were 396 complemented by time-resolved laser fluorescence spectroscopy (TRLF) measurements 397 that showed five water molecules surrounding the metal cations, which was interpreted 398 as a bidentate inner sphere complex.²² The similar distance to the second shell found by 399 Dardenne et al.²² and by us in these two different systems can be explained by the 400 smaller size of Lu in comparison to Y, which is compensated for by the longer Fe-O 401 bond distance in comparison to the Al-O bond, making possible a bidentate 402 mononuclear sphere at 3.4 Å in Fe oxides and Al oxides. Another study by Estes et al. 403 reported Eu K-edge EXAFS data of Eu(III) sorbed onto hematite.¹⁹ They observed a 404 decrease in the coordination number of the Eu-O bond from 8 to 5 upon adsorption 405 accompanied by a smaller Eu-Fe bond distance at 3 Å. This result was interpreted via 406 molecular simulations as being due to the formation of an inner sphere bidentate 407

408 mononuclear surface complex. Such a decrease in the coordination number was not 409 observed in this study; the CN of the Y-O correlation was maintained at $\sim 8 \pm 1$ for both 410 the aqueous and surface complexes.

411 Identification of Y species in waste samples from AMD treatment

The k³-weighted EXAFS of the column samples (upper part) and standards (lower part) 412 and their Fourier transform (FT) envelope functions with corrected phase-shifts are 413 414 shown in Fig. 3. Visual comparisons of the EXAFS spectra of the column samples reveal a strong similarity between them, suggesting similar local order of Y in all 415 416 samples. In addition, the EXAFS spectra of the column samples are very similar to solid references (B-YSO4-ads, B-Y-ads, B-Y-cop and A-Y-ads). They have a shoulder at 417 ~6.5 Å⁻¹, indicating a higher frequency which is absent in the aqueous samples (see 418 419 arrow in Fig. 3).

Based on the PCA analyses (see Chapter S3), the reconstructed spectra indicated that 420 421 only two components were required to describe the samples (Fig. S4). The two most 422 likely Y species identified by target transformation of the standard spectra, were the B-YSO4-ads and B-Y-ads references (Fig. S5). However, since their EXAFS signals are 423 424 very similar, LCF of the EXAFS spectra of the waste samples from column treatments was performed using B-YSO4-ads and YSO4-sol (Fig. 3A, upper part) to quantify the 425 426 proportions of different Y species in the waste samples. The use of these two standards would correspond to YSO_4^+ adsorbed forming inner- and outer-sphere complexes, 427 respectively. 428

The LCF results show that more than 75% of Y adopts a local geometry similar to that of the B-YSO4-ads reference (with the exception of sample W-Alm-C3-8, for which the LCF indicated equal proportion of both YSO4 presented as inner sphere and outer-

sphere complexes (Table 3)). The waste samples were selected from the depth where the 432 major proportion of Al precipitated, assuming basaluminite precipitation. However, 433 other minor mineral species may have also been precipitated, such as Fe-434 oxyhydroxisulfate, due to the presence of small amount of Fe in the solids. Specifically, 435 samples from Almagrera column (W-Alm-C3-8 and W-Alm-C3-9) presented higher Fe 436 concentration than samples from Monte Romero treatment. REE can be also scavenged 437 by amorphous Fe hydroxides in AMD neutralization environments,⁴⁸ so a minor or 438 439 almost negligible fraction of Y could be retained by the small proportion of ferric oxides precipitated in the selected samples, explaining the lower proportion of B-YSO4-440 441 ads.

442 Table 3. Percentages of Y species in the column samples obtained from LCF of the 443 EXAFS spectra. R-factor and χ^2 are defined in Chapter S3 of the SI.

Standard	C1-4	C1-5	C3-8	C3-9	
B-YSO4ads	0.719 ± 0.036	0.867 ± 0.038	0.495 ± 0.059	0.746 ± 0.061	
YSO4-solution	0.281 ± 0.069	0.133 ± 0.071	0.505 ± 0.084	0.254 ± 0.085	
R-factor	0.14	0.14	0.29	0.30	
χ2	133	154	371	392	

444

445 ENVIRONMENTAL IMPLICATIONS

446 The YSO₄⁺ aqueous species has been characterized combining PDF analyses of aqueous
447 solutions and AIMD simulations, confirming the formation of an inner-sphere Y-SO₄

ion pair with a monodentate configuration, with an Y-S interatomic distance of 3.5 Å. 448 Results from thermodynamic sorption model describe REE sorption onto basaluminite 449 via sorption of aqueous REESO_4^+ . The use of an atomistic model using this positively 450 charged ion yields the best results for the EXAFS fitting of Y sorbed on basaluminite. 451 However, the EXAFS technique cannot confirm the presence of YSO_4^+ sorbed into 452 basaluminite by itself, due to the low sensitivity to discern between Al and S neighbors. 453 454 Rather than this, the EXAFS fitting, together with the PDF, can confirm the strong 455 interaction and the formation of inner-sphere surface complexes of Y onto basaluminite precipitates, via ligand exchange with AlO₆ units of its structure. EXAFS analyses of 456 column waste samples show that most of the Y is retained as the same inner-sphere 457 sorbed species, YSO_4^+ , with a low proportion of YSO4+ in outer-sphere configuration. 458

The description of the local structure of yttrium sorbed onto the basaluminite surface 459 460 provided here complements the atomic configuration studies of other trace metals, such as As and Se oxyanions.⁴⁴ The chemical similarity between yttrium and other HREE 461 (from Tb to Lu) suggests that similar environments could be present for the other 462 463 elements of the same group. This fact has important environmental consequences, as the 464 HREE would be strongly sorbed, via covalent bonds, which could result in their long-465 term immobilization at least until the host phase is dissolved or re-precipitated. A key 466 question emerges about the long-term stability of the complex, particularly with an increase in the solution pH. The structural evolution of pure basaluminite with 467 468 increasing pH showed its recrystallization onto boehmite, releasing sulfate into the solution.⁴⁵ Previous studies on poorly crystalline Fe phases from AMD such as 469 schwertmannite have shown a release of adsorbed toxic elements during aging due to 470 recrystallization processes.⁴⁹ The occurrence of such processes in REE-doped 471 basaluminite could result in the remobilization of REEs and of other potentially 472

473 adsorbed toxic elements. The extent to which these processes can occur in natural474 waters requires further investigation.

Further studies on the potential uptake of LREEs by basaluminite are necessary to confirm and/or compare different structural mechanisms to those of their HREE counterpart, as an step to develop separation methods for REEs. These investigations will help to enhance a selective recovery of REE from AMD wastes and thus to generate an economic value of the waste generated during the neutralization of AMDaffected waters.

481 Associated information

Supporting information. The supporting information contains 1) five chapters with detailed description of synthesis of reference materials, analytical procedures and aqueous speciation calculations, synchrotron experiments and data analyses, molecular dynamics simulations and the Phyton code used to generate PDF from AIMD simulations; 2) three 3 tables describing samples used for EXAFS measurements and the modeling parameters from EXAFS fitting; and 3) four figures illustrating fitting of EXAFS spectra.

489 Acknowledgements

This work was funded by the European EIT 'Morerecovery' and the Spanish SCYRE (CGL2016-468 78783-C2-R)] projects. A.L. was also funded by a FPI grant (BES-2014-069978) and the "Make Our Planet Great Again" short-stay program. A.F.-M. acknowledges a grant from Labex OSUG@2020 (investissements d'avenir, ANR10 LABX56) for financial support of this research. A.L. thanks J. Bellés, M. Cabañas and R. Bartrolí (IDAEA-CSIC) for their analytical assistance. D.D.T. is grateful to the UK

496	Materials and Molecular Modelling Hub for computational resources, which is partially
497	funded by EPSRC (EP/P020194/1). This research utilised Queen Mary's Apocrita HPC
498	facility, supported by QMUL Research-IT. The authors would like to thank the ESRF
499	for in-house beamtime allocation at ID31 and the SOLEIL review committee for
500	beamtime allocation at BM30B-FAME on the French quota, proposal n. 30-02-1124.

501 **References**

- 502 (1) Binnemans, K.; Jones, P. T.; Blanpain, B.; Van Gerven, T.;, Yang, Y.; Walton, A.;
- 503 Buchert, M. Recycling of rare earths: A critical review. J. Clean. Prod. 2013, 51, 1–22.
- 504 https://doi.org/10.1016/j.jclepro.2012.12.037
- 505 (2) Kumar, R.; Jiang, C.; Wang, J.; Cietek, D.; Roth, J.; Gell, M.; Jordan, E. H. (2018).
- 506 Low Thermal Conductivity Yttrium Aluminum Garnet Thermal Barrier Coatings Made
- 507 by the Solution Precursor Plasma Spray: Part II—Planar Pore Formation and CMAS
- 508 Resistance. J. Therm. Spray Techn. 2018, 27(5), 794–808.
 509 https://doi.org/10.1007/s11666-018-0727-x
- 510 (3) Noack, C. W.; Dzombak, D. A.; Karamalidis, A. K. Rare Earth Element
- 511 Distributions and Trends in Natural Waters with a Focus on Groundwater. *Environ. Sci.*
- 512 *Technol.* **2014**, *48* (8), 4317–4326. https://doi.org/10.1021/es4053895.
- 513 (4) Evangelou V. P.; Zhang, Y. L. A Review: Pyrite Oxidation Mechanisms and Acid
- 514 Mine Drainage Prevention. Crit. Rev. Environ. Sci. Technol. 1995, 25 (2), 141–199.
- 515 https://doi.org/10.1080/10643389509388477.
- 516 (5) Coulton, R.; Bullen, C.; and Hallet, C. The design optimisation of active mine water
- treatment plants. *Land Contam.Reclamat.* **2003**, *11*, 273-280.
- 518 (6) Ackman, T. Sludge disposal from acid mine drainage treatment. U.S. Bureau of
- 519 Mines, *Report of Invest.* 8672, **1982**. Pittsburg, PA. 38 pp.
- 520 (7) Viadero, R. C.; Wei, X.; Buzby, K. M. Characterization and Dewatering Evaluation
- of Acid Mine Drainage Sludge from Ammonia Neutralization. *Environ. Eng. Sci.* 2006,
- 522 23 (4), 734–743. https://doi.org/10.1089/ees.2006.23.734.

- 523 (8) Hedin, R. S.; Watzlaf, G. R.; Nairn, R. W. Passive Treatment of Acid Mine 524 Drainage with Limestone. *J. Environ. Qual.* **1994**, *23* (6), 1338–1345.
- 525 https://doi.org/10.2134/jeq1994.00472425002300060030x.
- 526 (9) Younger, P. L.; Banwart, S. A.; Hedin, R. S. Mine Water Hydrology, Pollution,
- 527 *Remediation* 2002. https://doi.org/10.1007/978-94-010-0610-1.
- (10) Cravotta, C. A. Size and performance of anoxic limestone drains to neutralize
 acidic mine drainage. *J. Environ. Qual.* 2003, *32*, 1277–1289.
- 530 (11) Watzlaf, G. R.; Schroeder, K. T.; Kleinmann, R. L. P.; Kairies, C. L.; Nairn, R. W.
- The Passive Treatment of Coal Mine Drainage; laboratory report DOE/NETL2004/1202; U.S. Department of Energy, National Energy Technology Laboratory:
 Pittsburgh, PA, 2004.
- (12) Ayora, C.; Macías, F.; Torres, E.; Lozano, A.; Carrero, S.; Nieto, J.-M.; Pérez-534 López, R.; Fernández-Martínez, A.; Castillo-Michel, H. Recovery of Rare Earth 535 Elements and Yttrium from Passive-Remediation Systems of Acid Mine Drainage. 536 Environ. Sci. Technol. 2016. 50 (15)8255-8262. 537 538 https://doi.org/10.1021/acs.est.6b02084.
- (13) Ayora, C.; Caraballo, M. A.; Macias, F.; Rötting, T. S.; Carrera, J.; Nieto, J. M.
 Acid Mine Drainage in the Iberian Pyrite Belt: 2. Lessons Learned from Recent Passive
- 541 Remediation Experiences. Environ. Sci. Pollut. Res. 2013, 20 (11), 7837–7853.
- 542 <u>https://doi.org/10.1007/s11356-013-1479-2</u>
- 543 (14) Bigham, J. M.; Nordstrom, D K. Iron and aluminum hydroxysulfate minerals from
- acid sulfate waters. 2000 In: Jambor, J.L., Alpers, C.N., Nordstrom, D.K. (Eds.), Sulfate
- 545 Minerals, Crystallography, Geochemistry and Environmental Significance, vol. 40.

546 Mineralogical Society of America Reviews in Mineralogy and Geochemistry, pp. 351–
547 403.

- 548 (15) Carrero, S.; Fernandez-Martinez, A.; Pérez-López, R.; Lee, D.; Aquilanti, G.;
- 549 Poulain, A.; Lozano, A.; Nieto, J.-M. The Nanocrystalline Structure of Basaluminite, an
- Aluminum Hydroxide Sulfate from Acid Mine Drainage. Am. Mineral. 2017, 102 (12),
- 551 2381–2389. <u>https://doi.org/10.2138/am-2017-6059</u>.
- 552 (16) Gammons, C. H.; Wood, S. A.; Pedrozo, F.; Varekamp, J. C.; Nelson, B. J.; Shope,
- 553 C. L.; Baffico, G. Hydrogeochemistry and Rare Earth Element Behavior in a
- 554 Volcanically Acidified Watershed in Patagonia, Argentina. Chem. Geol. 2005, 222 (3-
- 4), 249–267. <u>https://doi.org/10.1016/j.chemgeo.2005.06.002</u>.
- (17) Lozano, A.; Ayora, C.; Fernández-Martínez, A. Sorption of rare earth elements
 onto basaluminite: the role of sulfate and pH. *Geochim. Cosmochim. Acta* 2019. 258,
 50-62.
- 559 (18) Gimeno Serrano, M. J.; Auqué Sanz, L. F.; Nordstrom, D. K. REE Speciation in
- 560 Low-Temperature Acidic Waters and the Competitive Effects of Aluminum. Chem.
- 561 Geol. 2000, 165, 167–180. https://doi.org/10.1016/S0009-2541(99)00166-7(19) Estes,
- 562 S. L.; Arai, Y.; Becker, U.; Fernando, S.; Yuan, K.; Ewing, R. C.; Zhang, J.; Shibata, T.;
- Powell, B. A. A Self-Consistent Model Describing the Thermodynamics of Eu(III)
 Adsorption onto Hematite. *Geochim. Cosmochim. Acta* 2013, *122*, 430–447.
 https://doi.org/10.1016/j.gca.2013.08.023.
- 566 (20) Yang, S.; Sheng, G.; Montavon, G.; Guo, Z.; Tan, X.; Grambow, B; Wang, X. 567 Investigation of Eu(III) immobilization on γ -Al₂O₃ surfaces by combining batch

technique and EXAFS analyses: Role of contact time and humic acid. *Geochim. Cosmochim. Acta* 2013, *121*, 84-104.

570 (21) Fan, Q. H.; Tan, X. L.; Li, J. X.; Wang, X. K.; Wu, W. S.; Montavon, G. Sorption

of Eu(III) on Attapulgite Studied by Batch, XPS, and EXAFS Techniques. *Environ. Sci.*

572 *Technol.* **2009**, *43* (15), 5776–5782. https://doi.org/10.1021/es901241f..

- 573 (22) Dardenne, B. K.; Schäfer, T.; Denecke, M. A.; Rothe, J.; Kim, J. I. Identification
- and Characterization of Sorbed Lutetium Species on 2-Line Ferrihydrite by Sorption

575 Data Modeling, TRLFS and EXAFS. *Radiochim. Acta* **2001**, *89*, 469–479.

- 576 (23) Allen, P. G.; Bucher, J. J.; Shuh, D. K.; Edelstein, N. M.; Craig, I. Coordination
- 577 Chemistry of Trivalent Lanthanide and Actinide Ions in Dilute and Concentrated

578 Chloride Solutions FT Magnitude. *Inorg. Chem.* **2000**, *39* (11), 595–601.

- (24) Lindqvist-Reis, P.; Lambe, K.; Pattanaik, S.; Persson, I.; Sandström, M. Hydration 579 of the Yttrium (III) Ion in Aqueous Solution . An X-Ray Diffraction and XAFS 580 Structural Chem. 104, 402-408. 581 Study. J. Phys. В 2000, https://doi.org/10.1021/jp992101t. 582
- 583 (25) Yaita, T.; Narita, H.; Suzuki, S.; Tachimori, S.; Motohashi, H.; Shiwaku, H. Structural Study of Lanthanides(III) in Aqueous Nitrate and Chloride Solutions by 584 EXAFS. J. Radioanal. Nucl. Chem. 1999, 239 371-375. 585 (2),https://doi.org/10.1007/BF02349514. 586

(26) Duvail, M.; Ruas, A.; Venault, L.; Moisy, P.; Guilbaud, P. Molecular Dynamics
Studies of Concentrated Binary Aqueous Solutions of Lanthanide Salts: Structures and
Exchange Dynamics. *Inorg. Chem.* 2010, 49 (2), 519–530.
https://doi.org/10.1021/ic9017085.

- 591 (27) Kowall, T.; Foglia, F.; Helm, L.; Merbach, A. E. Molecular Dynamics Simulation
- 592 Study of Lanthanide Ions Ln^{3+} in Aqueous Solution Including Water Polarization.
- 593 Change in Coordination Number from 9 to 8 along the Series. J. Am. Chem. Soc. 1995,
- 594 *117* (13), 3790–3799. https://doi.org/10.1021/ja00118a015.
- 595 (28) Ohta, A.; Kagi, H.; Tsuno, H.; Nomura, M.; Kawabe, I. Influence of Multi-Electron
- 596 Excitation on EXAFS Spectroscopy of Trivalent Rare-Earth Ions and Elucidation of
- 597 Change in Hydration Number through the Series. Am. Mineral. 2008, 93 (8–9), 1384–
- 598 1392. https://doi.org/10.2138/am.2008.2628.
- (29) Rizkalla, E. N., Choppin, G. R. Lanthanides and Actinides Hydration and
 Hydrolysis. In: Handbook on the Physics and Chemistry of Rare Earths. 1994.
 (Gschneider, K. A., Eyring, L., Choppin, G. R., Lander, G. H., eds.) Vol. 18 –
 Lanthanides/Actinides: Chemistry. Elsevier Science B.V., Amsterdam, The
 Netherlands.
- (30) Duvail, M.; Spezia, R.; Vitorge, P. A Dynamic Model to Explain Hydration
 Behaviour along the Lanthanide Series. *Chemphyschem* 2008, 9, 693–696.
 <u>https://doi.org/10.1002/cphc.200700803</u>.
- 607 (31) Janicki, R.; Starynowicz, P.; Mondry, A. Lanthanide carbonates. *Eur. J. Inorg.*608 *Chem.* 2011, 24, 3601–3616. https://doi.org/10.1002/ejic.201100184.
- 609 (32) Jeanvoine, Y.; Miró, P.; Martelli, F.; Cramer, C. J.; Spezia, R. Electronic structure
- and bonding of lanthanoid(III) carbonates. Phys. Chem. Chem. Phys. 2012, 14 (43),
- 611 14822–14831. <u>https://doi.org/10.1039/c2cp41996c</u>.
- 612 (33) Harris, S. M.; Nguyen, J. T.; Pailloux, S. L.; Mansergh, J. P.; Dresel, M. J.;
- 613 Swanholm, T. B.; Gao, T.; Pierre, V. C. Gadolinium complex for the catch and release

- 614 of phosphate from water. Environ. Sci. Technol. 2017, 51 (8), 4549–4558.
 615 https://doi.org/10.1021/acs.est.6b05815.
- 616 (34) Luo, Y. R.; Byrne, R. H. Carbonate complexation of yttrium and the rare earth
- elements in natural rivers. *Geochim. Cosmochim. Acta* **2004**, 68, 691–699.
- 618 (35) Torres, E.; Auleda, M. A Sequential Extraction Procedure for Sediments Affected
- 619 by Acid Mine Drainage. J. Geochemical Explor. 2013, 128, 35–41.
- 620 (36) Parkhurst, D. L.; Appelo, C. A. J. User's guide to PhreeqC (version 2.18) A
- 621 computer program for speciation, and inverse geochemical calculations, U.S.
 622 Department of the Interior, U.S. Geological Survey 1999.
- (37) Blanc, P.; Lassin, A.; Piantone, P.; Azaroual, M.; Jacquemet, N.; Fabbri, A.;
 Gaucher, E. C. Thermoddem: A Geochemical Database Focused on Low Temperature
 Water/Rock Interactions and Waste Materials. *Appl. Geochemistry* 2012, 27 (10), 2107–
- 626 2116. https://doi.org/10.1016/j.apgeochem.2012.06.002.
- (38) Juhás, P.; Davis, T.; Farrow, C. L.; Billinge, S. J. L. PDFgetX3: a rapid and highly
 automatable program for processing powder diffraction data into total scattering pair
 distribution functions. *J. Appl. Crystallogr.* 2013, 46, 560–566.
 https://doi.org/10.1107/S0021889813005190.
- 631 (39) Proux, O.; Biquard, X.; Lahera, E.; Menthonnex, J. J.; Prat, A.; Ulrich, O.; Soldo,
- 632 Y.; Trévisson, P.; Kapoujyan, G.; Perroux, G.; Taunier, P.; Grand, D.; Jeantet, P.;
- 633 Deleglise, M; Roux, J-P. and Hazemann, J-L. FAME A New Beamline for XRay
- 634 Absorption Investigations of Very Diluted Systems of Environmental, Material and
- 635
 Biological
 Interests.
 Phys.
 Scr.
 2005,
 970-973.
- 636 https://doi.org/10.1238/physica.topical.115a00970.

- 637 (40) Hutter, Y.; Iannuzzi, M.; Schiffmann, F.; VandeVondele, J. CP2K: Atomistic
- 638 Simulations of Condensed Matter Systems. <u>Wiley Interdiscip.</u> Rev. Comput. Mol. Sci.
 639 2014, 4, 15–25.
- 640 (41) Zhang, Y.; Yang, W. Comment on 'Generalized Gradient Approximation Made
 641 Simple". *Phys. Rev. Lett.* **1998**, *80*, 890–890.
- (42) Grimme, S.; Antony, J.; Ehrlich, S.; Krieg, H. A Consistent and Accurate Ab Initio 642 643 Parametrization of Density Functional Dispersion Correction (DFT-D) for the 94 H-Pu. J. 2010, 644 Elements Chem. Phys. 132 (15),154104. 645 https://doi.org/10.1063/1.3382344.
- (43) Juhás, P.; Farrow, C. L.; Yang, X.; Knox, K. R.; Billinge, S. J. L. Complex 646 Modeling: A Strategy and Software Program for Combining Multiple Information 647 Sources to Solve III Posed Structure and Nanostructure Inverse Problems. Acta 648 649 Crystallogr. Adv. 2015, A71. 562-568. Sect. A Found. https://doi.org/10.1107/S2053273315014473. 650
- (44) Carrero, S.; Fernandez-Martinez, A.; Pérez-López, R.; Poulain, A.; Salas-Colera,
 E.; Nieto, J. M. Arsenate and Selenate Scavenging by Basaluminite: Insights into the
 Reactivity of Aluminum Phases in Acid Mine Drainage. *Environ. Sci. Technol.* 2017, *51*(1), 28–37. https://doi.org/10.1021/acs.est.6b03315.
- (45) Lozano, A.; Fernández-Martínez, A.; Ayora, C.; Poulain, A. Local Structure and
 Ageing of Basaluminite at Different pH Values and Sulphate Concentrations. *Chem. Geol.* 2018, 496, 25–33. https://doi.org/10.1016/j.chemgeo.2018.08.002.
- 658 (46) Cruz-Hernández, P.; Carrero, S.; Pérez-López, R.; Fernández-Martínez, A.;
- 659 Lindsay, M. B .J.; Dejoie, C.; Nieto, J. M. Influence of As(V) on precipitation and

- transformation of schwertmannite in acid mine drainage-impacted waters. *Eur. J. Mineral.* 2018 https://doi.org/10.1127/ejm/2019/0031-2821.
- 662 (47) Rabung, Th.; Geckeis, H.; Wang, X. K.; Rothe, J.; Denecke, M. A.; Klenze, R.;
- 663 Fanghäel, Th. Cm(III) sorption onto γ -Al₂O₃: New insight into sorption mechanisms by
- time-resolved laser fluorescence spectroscopy and extended X-ray absorption fine
- 665 structure. *Radiochim. Acta* **2006**, *94*, 609-618.
- 666 (48) Verplanck, P. L.; Nordstrom, D. K.; Taylor, H. E.; Kimball, B. A. Rare earth
- 667 element partitioning between hydrous ferric oxides and acid mine water during iron
- 668 oxidation. *Appl. Geochem.* **2004**, *19*, 1339–1354.
- 669 (49) Acero, P.; Ayora, C.; Torrentó, C.; Nieto, J. M. The Behavior of Trace Elements
- 670 during Schwertmannite Precipitation and Subsequent Transformation into Goethite and
- 671 Jarosite. *Geochim. Cosmochim. Acta* **2006**, *70* (16), 4130–4139.